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ISSN=27769

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URBAN AND REGIONAL DEVELOPMENT CENTER
UNIVERSITY OF FLORIDA

GAINESVILLE, FLORIDA

74-74-03

This paper has now been published
by the Center for Wetland
Resources, La. State Univ.,
Baton Rouge, LSU-SG-74-03

WORK PAPER NO. 3

THE VALUE OF THE TIDAL MARSH

BY

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MAY 1, 1973

Pre-publication draft subject to revision.
Comments appreciated by Dr. Eugene P. Odum.

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- the tendency for the value of the fish resource to be driven to zero in a market context is an example of market *failure* arising from common property problems in the fishery. That is, the zero market value on the fish resource is not condoned by conventional economics as GOP imply, but is cited as a classic example of market failure (Haveman, 1973).
13. One can see the possible fallacy of this assumption by reference to an example. Assume that an acre of wetland can produce a ton of marine worms per year. Furthermore, assume that a ton of marine worms could be artificially propagated in a laboratory at a cost of \$100,000. Could we then conclude that wetland services that produce a ton of marine worms are worth \$100,000 to society? The answer is an unequivocal no, unless we can convincingly demonstrate that society would be willing to pay \$100,000 per ton for marine worms.
 14. One particular waste assimilation benefit of marshland ignored by GOP, and by this discussion, is the assimilation of waste from nonpoint discharges. The cost of control of these wastes may be quite high, and the alternative of natural wetlands as a treatment source may be quite valuable.
 15. The discussion in this article has not exhausted the various methodological issues that cast doubt on the validity of the GOP estimates. Other issues include their (1) failure to properly distinguish among marginal, average, and total values of wetlands, (2) choice of the appropriate discount rate for calculating present values, and (3) neglect of tradeoff possibilities between wetland services.

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THE VALUE OF THE TIDAL MARSH

by
James G. Gosselink¹, Eugene P. Odum², and R. M. Pope³

The objective evaluation of different land use strategies has been severely hampered by the difficulty of stating the value of different objectives in a common currency. Cost accounting techniques for industrial, commercial and residential development are well developed and these interests can bring strong pressures to bear because of the nearly universal acceptance of evaluation techniques which show the cash value of a particular management alternative. Against these evaluation techniques conservationists and natural resources economists have been at a disadvantage because of the difficulty of translating the value of natural or undeveloped areas into monetary terms. Frequently, therefore, the alternative management decision of leaving land in its natural state is not adequately defended or seriously considered. Although recreation, for instance, is recognized as a legitimate land use it is difficult to place a cash value on the esthetic pleasure derived from an unspoiled forest or a natural lake (for a discussion of this problem, see Pope, 1972). As a result, hearings on proposed land use developments are usually charged with a great deal of emotion and frustration for all parties involved. In this paper we develop a step-wise means of assessing

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the true value of natural tidal marshes to society as a whole - a value based, not only on commercial usage, but on social usage and the moral value of natural (i.e. "undeveloped") estuarine environment.

Tidal marshes are lands which are particularly vulnerable to capricious development (W. E. Odum, 1970), because many of the real values of marshes are not recognized, or accrue some distance from the marsh itself. Teal (1962) estimated that 45% of the net primary production of a Georgia Spartina alterniflora marsh was flushed into the adjacent bay by tidal action. Odum and de la Cruz (1967) estimated the net export of organic matter (which includes many mineral nutrients) from 25 hectares of such marsh was 40 kgms and 140 kgms on a neap and spring tidal cycle respectively. Stowe et al., (1971) have estimated that well over one-half of the total production of organic matter in the Gulf Coast estuary originates from the surrounding marshes. In this and other coastal marshes and other shallow water production areas (reefs, sea grass beds, etc.) all over the world export mineral and organic nutrients that support much of the production of the adjacent estuarine and coastal waters (see E. P. Odum, 1971). Furthermore, as is well documented, estuaries serve as a nursery ground for commercially important coastal fish and shellfish. McHugh (1966) estimates that two-thirds of the cash value of species harvested on the Atlantic and Gulf Coasts is "estuarine dependent". Thus, productive marshes are an integral part of the estuarine system, which not only exports nutrients but also grows sea food that may be harvested in adjacent waters. Nursery ground is the only valuable function of an undisturbed marsh, but it is an important

value and now generally recognized one. Even though the marsh may be privately owned the production of that marsh does not, at present, accrue directly to the owner, but to a commercial fishery, perhaps many miles away. Thus, the true value of a flowing-water, exporting system must be based on a much broader cost-accounting than is usually employed in real estate evaluation.

At the other end of the spectrum of interested parties is the land developer. Since the coasts are often the most densely populated parts of the country, there is strong pressure to fill marshland for commercial, industrial and residential use. California has lost 67% of its marshland in this way; New York and New Jersey 10-25% (Sweet, 1971). Additionally, due to alterations in natural processes brought about by land management decisions, Louisiana is currently losing 16.5 square miles of marshland per year (Gagliano, Kwon and van Beek, 1970). Indeed, with high marsh in prime commercial areas of New Jersey selling for as much as \$80,000 per acre the economic incentive to develop is extremely strong.

In the paragraphs below we present first the value of a natural tidal marsh-estuary based on identifiable present commercial and recreational uses for which monetary values can rather well be determined. Since this omits a number of other values the marsh has, which are identifiable but more difficult to quantify, we discuss next potential additional values and attempt to equate these with dollars. Thirdly, we calculate the total "life support" value of the tidal marsh according to the procedure suggested by Odum and Odum (1972). Finally, a summary table is presented as a method of integrating and summing values as a basis for land use planning in the coastal zone.

Fishery Production Based on Harvest of Naturally Produced Animals (Natural Secondary Production)

Since, as already acknowledged, fishery production in estuarine coastal waters is linked to shallow water production zones such as marshes, reefs and sea grass beds, one can estimate the present value of an acre of marsh and its associated tidal creeks by evaluating the dependent commercial fishery. For instance, on the Georgia coast the dockside value of fish and shellfish (including shrimp) in 1965 was 3.66 million dollars (Carley, 1968). Value added in processing amounted 5.23 million dollars raising the total value to 8.9 million (Table 1). Georgia has 393,000 acres of coastal marshland (Spinner, 1969). Directly proportioned this works out to about \$23 per acre per year. Sport fishing along the Georgia coast is estimated to involve 280,000 fishermen who spend an average of \$80 each per year (E. P. Odum, 1968). Other recreational uses such as hunting and boating are arbitrarily valued at one-half the sport fishing so that a total "sports" value comes to \$108 per acre per year value (Table 1). Comparable statistics for Louisiana and Florida marshes are shown in Table 2. They vary somewhat from the Georgia figure but yield similar estimates suggesting a minimal value of about \$100 per year per acre of marsh just from the standpoint of fishery and recreational values.

These figures place a value upon a piece of real estate which is easily comprehended in terms of present evaluation techniques. Using the income-capitalization approach (see Barlowe, 1965, page 180) and the formula, $V = R/i$ where V represents the value of a parcel of land, R represents the annual return from it, and i represents the appropriate

(i.e., interest rate, it may be easily seen that the minimum value of an acre of marshland, due to fishery and recreational returns is \$2000, if $R = \$100$ and we assume an interest rate of 5%.

Admittedly the estimate of \$2000 can be questioned since one cannot prove that all the fishery would be destroyed if the marshes were; nor can one say exactly how many acres of marsh are necessary to support the present level of fishery activity in estuarine and offshore waters. The low Spartina alterniflora marshes, because they are subjected to vigorous tidal flushing that acts as an "energy subsidy", are certainly more valuable per unit area for estuarine productivity than the higher, dwarf Spartina or Juncus marshes (Odum and Fanning, 1973). Of course, if the latter cover large areas, the total contribution can be considerable. As we aim to show even an inflated value based only on the harvest of naturally produced by-products falls far short of the value obtained by a more complete cost accounting that includes other, and in the long run, more important considerations.

Aquaculture Based on the Utilization of Natural Primary Production

Another approach to evaluation of land is to consider its potential for development, and this is usually the major factor in conventional real estate dealings. In estuaries, development could range anywhere from intensive aquaculture to draining and filling for industrial or other use. Since this paper is concerned with evaluation of natural marshes we will consider only practices which would use the marsh as a renewable resource but retain it in its more or less self-maintaining natural state. Oyster aquaculture provides a promising possibility since less modification of

the estuary is required than for intensive shrimp or fish culture. Coastal estuaries can certainly support oyster production on a more intensive scale than is found at present. Estimates for income that might be obtained from this kind of development are shown in Table 3. Annual yields of 1800 pounds of oyster meat per acre worth \$100 (value added) were obtained by the late Dr. Robert Lunz, with moderate intensive culture in the marsh-bordered estuaries at Bear Bluff (E. P. Odum, 1968). Very intensive raft culture, as developed with rafts covering 1/10 the water surface, could theoretically yield as high as 4500 pounds of oyster meat per acre with a value of \$260 per year (Table 3, column 1). This kind of intensive aquaculture is possible only in flowing water systems where the organic production of algae passes across the oyster rafts, and the feces produced by the oyster population are also carried away from the rafts.

In still another approach Stowe et al., (1970) estimated the annual primary production of estuarine waters in Louisiana to be about 18,000 pounds dry weight per acre. Half of this originates in the marsh and is flushed into estuarine waters by tides. Using these figures and assuming a conversion efficiency of 10% from the primary photosynthetic product to oyster meat, it would appear that intensive raft culture of oysters would require about three acres of marsh and three acres of estuary per acre of oyster raft. On the basis of an acre of marsh, then, the potential value from oyster culture, including value added during processing, would be about \$350 per year for a moderate culture level, and nearly \$900 per year with intensive raft culture (Table 3, Column 1).

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Waste Treatment Work as a Basis for Economic Evaluation

The shortcomings of evaluating environment only in terms of direct uses or products is that such cost-accounting ignores the extremely valuable life-support work natural areas carry out without any development or direct use by man. It is this "free work of nature" that is grossly undervalued, simply because it has always been taken for granted, or assumed to be unlimited in capacity. Since development by man may adversely affect this work, it is important to evaluate it before deciding on what kind of development, if any, is in the long term best interest of both the environment and the economy.

One very important contribution estuaries make to the growth and economic wealth of highly urbanized regions is the waste treatment that active ecosystems can accomplish without appreciable reduction in water quality. Sweet's (1971) estimate of the waste assimilated by five mid-Atlantic estuaries, in terms of BOD load received, is shown in Table 4. In general, the sewage discharge in these estuaries does not stop at the "secondary" stage but continues through the "tertiary" stage of nutrient removal and assimilation. Since artificial tertiary treatment of sewage is very much more expensive than secondary, then an acre of marsh-estuary is doing about \$14,000 worth of work per year at a daily loading of 19.4 lb. BOD, assuming the cost of artificial tertiary treatment is at least \$2/lb BOD (Table 5). In other words, this is what it would cost man to deal directly with his wastes if the acre was not available to do this work. Resorting, again, to the income-capitalization calculation an acre of estuary that is able to handle the mean waste loading shown in Table 4 is worth a whopping \$280,000 (Table 5, Column 2). It is no wonder that

large cities and industrial complexes tend to be located where bodies of water are available for "free" treatment plants!

Of course, it is apparent that mid-Atlantic estuaries are overloaded (see especially Hudson and James estuaries, Table 4) to such an extent that oxygen and other water quality aspects are reduced to an undesirable level, especially in terms of fisheries and recreation. A value of \$280,000/acre thus represents a large "overload" of work which has serious pollution side-effects, and if continued or increased will result in system breakdown. If the BOD load can be reduced, the estuaries would function better as tertiary treatment plants and are valuable overall.

In a detailed study of the Delaware estuary the Federal Water Pollution Control Administration estimated the cost of water treatment to increase the minimum dissolved oxygen level to 4.5 mg/litre (a minimum water quality level) at \$460 million (amortized capital and operating costs for 25 years). This works out to \$264 per acre per year. The resulting improvement in water quality would yield recreational benefits and, more important, reduce stress on the system's ability to do tertiary treatment. \$264/acre seems a small price to pay to insure that the acre can safely do thousands of dollars of work free!

It is clear, then, that estuaries are not really effective as secondary treatment since large amounts of organic matter introduced into systems naturally high in organic detritus reduces the dissolved oxygen levels to an undesirable extent. And, as we have seen, the economic value of estuaries as secondary treatment plants is relative.

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since the energy* and money necessary for artificial secondary treatment is not large per unit volume of waste. The most important contribution marshes and estuaries can make in waste treatment is in tertiary treatment to remove and recycle inorganic nutrients, a very expensive process if carried out by man in artificial systems as we have seen. When nutrient rich effluents enter a marsh-estuarine system the nutrients are effectively trapped by the tidal circulation pattern (Bowden, 1967), and assimilated in the productive biological system. Estuarine ecosystems have evolved adaptations to high nutrient levels, and have a large capacity to buffer nutrient changes. Pomeroy, et al., (1972) have shown that the phosphate recycle system is so large and homeostatic in Georgia estuarine and marsh sediments that the level of phosphate in those waters varies little throughout the year, despite variations in input. Studies in Louisiana (Ho et al., 1970) confirm this. The sediments act as both source and sink, effectively buffering the effects of large additions of phosphate to the estuarine system.

Although research results are not as clear for nitrogen, flooded marshes appear to be uniquely adapted for denitrification and, therefore, may be extremely valuable for treatment of inorganic nitrogen wastes also. Studies in flooded swamp and marsh soils (Patrick et al., 1971) have shown substantial loss of inorganic nitrogen by denitrification in the anaerobic zone.

* A Rand report (R-1098-NSF; 1972) just seen lists the amount of electrical energy needed for secondary treatment of 10^6 gallons municipal wastes as 660 kWhr (about 56.8×10^4 Kcal) which on a per capital basis works out to be less than 1% of the electricity now consumed in an urban area.

Experimental confirmation of these important water quality of marsh-estuarine systems is slowly evolving. Valiela and Teal treated salt marsh plots with sludge from a secondary sewage treatment plant, and measured the inorganic nitrogen and phosphorous loss first tide following each application. From late May through mid-November they applied $25.2 \text{ g sludge m}^{-2} \text{ wk}^{-1}$ for a total of about 1000 pounds N and 400 pounds P were removed per acre of marsh in these

Grant and Patrick (1970) give a second example in Tinicum marsh, Pennsylvania. Water flowing out of this marsh showed an average reduction per acre of 6.4 pounds of phosphorous (as phosphate) and 100 pounds of nitrogen (nitrate and ammonia), as compared with pollution flowing into the marsh.

Using Culp's et al., (1966) estimates of tertiary treatment cost of \$100 per million gallons and Seibel's (1969) estimate of 10 ppm phosphorous in sewage water, 83 pounds phosphorous could be removed at a cost of \$100. At this rate, the work done by the marsh in phosphorous removal is worth \$480 per year (Valiela and Teal's data) to \$1420 per year (Grant and Patrick's data, assuming effective biological activity 185 days per year). Income-capitalization of these data yield a per acre value of \$9,600 and \$28,400, respectively, for removal of only one major

In summary, it is clear that man should pay for secondary treatment of wastes since such treatment is relatively inexpensive and untreated organic materials greatly stress any natural aquatic system, but especially marsh-estuaries. However, man will and should depend on primary natural ecosystems for tertiary treatment of huge volumes of low strength wastes which would be extremely expensive to treat artificially.

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the economic value of estuaries as tertiary treatment plants can be valued in tens of thousands of dollars per acre as compared to mere hundreds that accrue from by-product uses. The shallow-water zones occupied by marshes play a major role in this very valuable life support work since their contribution to the overall metabolism of the estuary is proportionally high.

Other Marsh Functions

Other functions of the natural marsh are more difficult to quantify, but no less real. Perhaps the most important of these is the role of the marsh in global cycles of nitrogen and sulfur. The continuing normal function of the biosphere depends on the chemical reduction of carbon, nitrogen and sulfur, which are incorporated into all living tissues. While carbon reduction occurs through photosynthesis in oxidizing atmosphere, completion of the cycle of the other two elements depends on microbial action in a reducing environment (Deevey, 1971). Nitrogen fixation in the world has been nearly doubled by industrial fertilizer production (Delwiche, 1970). Some of this reduced nitrogen is accumulating in the slowly increasing biomass on the earth's surface, as shown by eutrophied water bodies, but apparently the global biosphere has compensated for increased nitrogen inputs to some extent by increased denitrification. This microbial process requires the close proximity of oxidized and reduced zones. Nitrogen of biological origin is oxidized to nitrate in the oxidized layer, diffuses into the reduced zone and is reduced to nitrogen gas, escaping to the atmosphere. Tidal marshes are ideally suited for this function. Tidal waters carry nutrients to the

marsh surface where they diffuse through a thin layer of oxidized
iment to the anaerobic zone below. The sulfur cycle, in the same
depends on reduction of sulfates in anaerobic muds to sulfur and su
fides. Oxygen is a by-product of the reaction. In Linsley Pond, (C
sulfate reduction may be as much as 10% of carbon reduction (Deevey
1971), so its magnitude is of some significance. The industrial con
tribution to atmospheric sulfur has increased to about one-third of the
total atmospheric sulfur burden (Kellogg et al., 1972). This sulfur
washed from the atmosphere by rain, primarily as sulfate. The lack
widespread accumulation of sulfuric acid is evidence of the efficien
of the sulfate reduction system in anaerobic muds.

These two processes have not been quantified on an area basis in
estuarine systems but impressive evidence points to the importance of
the coastal anaerobic muds to continued normal functioning of global
cycles of nitrogen and sulfur.

There are still other marsh functions worthy of mention for which
cost accounting is yet to be accomplished. A salt marsh is an import
buffer against storms. In particular, it absorbs the enormous energy
storm waves and acts as a water reservoir for coastal storm waters, th
reducing damage further inland. Some idea of the protective value of
wide band of energy-absorbing marshes and barrier islands is seen in t
increasing national cost for "disaster relief" in coastal areas which
either lack these natural protective "breakwaters" or where they have
been filled in or "bulkheaded" for housing or other development; marsh
and island protected coasts suffer comparatively little damage even in
fierce hurricanes. Rising costs of coastal development is very often

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the result of ill-planned modification of natural protective systems, not the result of increased storm intensity.

It is also becoming apparent that marshes are important in the protection of the beautiful white sand beaches of the outer barrier islands and "banks" (see Hoyt, 1967; Dolan, Godfrey, and W. E. Odum, 1973). Where the energy and muddy sediments of storm tides can be absorbed by large areas of marsh-estuary the natural erosion of beaches is at least balanced by formation of new beaches. Beaches are degraded or lost where they receive the full brunt of storm tides unless man resorts to very expensive artificial breakwaters. Therefore, one is justified in adding some of the enormous economic value of outer beaches to the value of the inner marsh-estuaries. The powerful flow of water in and out of large tidal basins also tend to keep harbors and inlets "dredged" - another example of useful "free work of nature". Recently reprinted in "Benchmark Papers in Geology" (Coates, ed. 1972) is an article by M. Burrows (Page 350), published in 1888 that describes how all of the early harbors on the southeastern coast of England were silted in when the great marshes were first diked and filled in; constant dredging and "a vast expenditure of national funds" then became necessary to keep harbors operational.

One value of coastal marshes that is generally recognized is their importance as habitats for migratory birds which have esthetic and hunting values, not only locally, but elsewhere in the continent. Tidal marshes that receive large inputs of freshwater are especially valuable in this regard. For example, almost the entire North American population of snow and blue geese (millions of birds) are dependent on the marshes

of the Texas and Louisiana coasts, which are their sole winter grounds. Some of these same low salinity marshes are also known for their production of muskrat fur, but again the monetary value of such by-products on an acre basis is not large.

Life Support Value as a Function of Productive Energy Flow

So far in this article we have resorted to the "component" approach that is, identifying and separately evaluating products, uses, and functions that are judged to have a value, or potential value. The shortcomings of this approach lies in the difficulty of identifying or summing the component values, because many of the uses conflict with one another. Thus, intensive aquaculture would reduce sport fishing and recreational boating values, or heavy use for secondary treatment of sewage would greatly reduce many other values - and so on. Therefore, it is difficult to obtain an overall value by the component approach. Also, most of the component values so far discussed relate to the estuarine system and not to the marshes per se; yet it is the marshes that need to be valued in monetary terms since they are the part of the system most vulnerable to modification and development by man.

H. T. Odum (1971) has suggested an "ecosystem" approach to valuing the total work of nature into monetary terms, so that the value of a delimited natural area can be determined without having to specify how the work flow might be divided into different uses and functions. Odum and Odum (1972) have extended this approach to land-use planning in which natural areas are considered as part of man's total environment. Since the exchange of energy

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is the basis for economic transactions it is suggested that the ratio of Gross National Product to National Energy Consumption can be used to equate energy with money. In round figures for the United States 10^{16} Kilocalories are consumed yearly to produce a Gross National Product of 10^{12} dollars, so that approximately 10^4 Kilocalories is equal to one dollar. Since the rate of primary production is a measure of energy flow of a natural community and an index of the useful work that might be accomplished, the ratio can be used to place a dollar value on any part of the natural environment where primary production can be measured or estimated.

Several recent estimates of the annual net primary production of coastal salt marshes are listed in Table 6. Several of these estimates, and most of those published prior to 1968, are underestimates because they are based on "standing crops" uncorrected for dry matter exported by the tides during the annual cycle. We judge the Louisiana and Georgia figures, as shown in Table 6, to be most representative of the highly productive marshes of the Gulf and south Atlantic coasts. Since, as already indicated, productivity is to a certain extent a function of water flow separate estimates are given for the higher or inner marshes (that receive less water flow subsidy) and the outer or low marshes (Table 6, Columns 1 and 2, respectively). Conservative estimates in round figures, then, are: 1000 gm/M^2 for high marshes; 3000 gm/M^2 for well-irrigated low marshes; and 2000 gm/M^2 for large areas of total marsh with an approximate 50-50 distribution of high and low types. Satisfactory measurements of total, or gross, primary production have not yet been accomplished for salt marshes. Since *Spartina* grasses utilize

the recently discovered C_4 photosynthetic pathway, the amount of photosynthate dissipated in respiration is certainly not as large as earlier estimated by Teal (1962). Based on the efficiency of plants adding 25% to the net production would give a reasonable estimate of gross production. Thus, annual gross primary production is of the order of 2500 gm dry matter/ M^2 overall, 1250 gm/ M^2 for high marshes and 3750 gm/ M^2 for the more productive low marsh stands.

To estimate the dollar value of an acre of marsh based on the energy/money conversion outlined in the second paragraph of this report we need only to multiply the round figure productivity estimates to get Kcal/ M^2 (see Odum and Fanning, 1973), and by 4046 to get dollars/acre and divide by 10^4 to get dollars/acre. Such a calculation gives a value of \$4,147/year for the marsh as a whole (range: \$2,703 for high marsh and \$6,2207 for low marsh). The income-capitalized value would be \$82,940 per acre overall. This is a larger value than obtained by summing of the component estimates, except for the "overloaded" tertiary marsh value (Table 5), which as already discussed is unreasonably high and also one that involves very large volumes of river and bay water flowing with the marshes. The advantage of cost accounting based on productivity (i.e. capacity for life support work) is that it can be applied to a particular acre, or acres, of marshland itself as it functions as a part of the whole estuary.

Discussion

Round figure values based on by-products, waste treatment and productivity are summarized in Table 7. The values of estuaries for

waste assimilation and general life support is greater than that accruing from by-products, even intensive aquaculture which if carried out on a large scale would eliminate most recreational and other uses. Summing values for components that could conceivably be non-competitive gives a "multiple-use" value approaching that based on productivity. As already emphasized the latter value (Table 7, No. 5) pertains directly to the marshlands, whereas all the other values summarized in Table 7 are based on the estuarine system as it functions as a whole.

Demonstrating that marshlands and estuaries have a substantial dollar value in their natural state certainly provides a big boost to preservation of such areas that are in public ownership. If large values such as those in Table 7 (items 4 and 5) are generally recognized and accepted, then state or federal agencies or commissions which have jurisdiction over the property or resource will be less likely to lease, give away, or sell valuable marshlands for capricious development. Also, planners will have a greater incentive and public support for zoning such areas into permanent protective categories. However, if the marshlands is in private ownership, the owner will stand to gain by selling for development no matter how high the appraisal, since leaving the area in its natural state would earn the owner little or no return. The dichotomy of interests between the value to the owner and the value to society becomes an increasing serious problem as population growth and industrial development accelerates. The pricing system, which one school of economics holds will solve all economic problems if left to operate unhampered, offers no solution to this problem since development becomes

essentially an irreversible action. Thus, even though the value of marshland increases as it becomes scarcer to an eventual point where its life support value could "outbid" other land uses there is no way to convert the previous development back to its former (and now irrevocable) state. The irony of dependence on the price system is that one can make a reasonable sounding argument for developing marshland and one can even offer an argument that a point will be reached when the marsh should be converted back to marsh, but it cannot effectively be done. Converting marshland, a very expensive process, even if technically possible.

Evaluation of marshland as a renewable resource - e.g. as a stream stretching into the future and increasing continually - is one way to alleviate the destructive tendency inherent in the present system as it now operates. The time has come to seek ways to insure that owners of natural resources with value to society receive a return. Direct purchase by Government is one solution, of course; scenic easement and tax relief are other approaches. Setting up "banks" where the owner is paid not to develop (as in "soil bank") is perhaps a feasible "delayed option" procedure in cases where outright purchase can not be made at a particular time.

The best solution is a "look ahead" land-use plan which delineates the amount and location of life-support natural areas that will be necessary to support a future desirable level of development. Such areas can then be acquired or zoned into the public domain before the onset of land speculation raises the market price. Odum and Odum (1971) present an overall model to show how the ratio of undeveloped to developed compartments could be objectively determined. Since many coastal

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are more productive than adjacent areas, they would generally receive high priority for inclusion into the undeveloped compartment.

Summary

Four levels were selected for monetary evaluation of marshlands and estuaries of the South Atlantic and Gulf Coasts: (1) by-product production (fisheries, etc.); (2) potential for aquacultural development; (3) waste assimilation; and (4) total "life support" value in terms of the "work of nature" as a function of primary production. Money values of marsh-estuaries in their natural state were calculated in terms of (a) annual return and (b) an income-capitalized value. Round figure values per acre at the four levels were: (1) a, \$100; b, \$2,000; (2) a, \$350; b, \$7,000; (3) a, \$2,500; b, \$50,000; (4) a, \$4,150; b, \$83,000.

The value of waste assimilation and total life support work (levels 3 and 4) are at least an order of magnitude higher than that which can be obtained from by-products, even under intensive aquacultural development which in itself would eliminate recreational and most other uses. These high values (levels 3 and 4) represent estimates of what man would have to pay (i.e. "internalize") in terms of the value of the useful work of an acre of estuary should it not be available to do this work. Summing values for specific functions judged to be non-competitive results in a value approaching that obtained by a total life support calculation (level 4), but the weakness of such a "component approach" is that most "multiple uses" do, in fact, compete at high levels of use. The advantage of level 4 cost accounting is that it can be applied to a particular acre, or acres, of marshland without having to specify how the work flow might

be divided into different uses and functions (which will change from time to time and place to place).

Detailed analysis of waste assimilation shows that marshland estuaries are not very effective (and, therefore, not very suitable) for secondary treatment of municipal wastes, but that they have a tremendous capacity for tertiary treatment of nutrients, especially nitrogen and phosphorous. Since secondary treatment is relatively inexpensive and tertiary treatment very expensive if done by man in artificial systems, it is clear that the large BOD loading now borne by many estuaries should be greatly reduced by organic matter digestion in municipal treatment plants in order that the natural systems can effectively carry out tertiary treatment and maintain a water quality that preserves and increases sea food production, recreation and other by-products.

Demonstrating that marshlands have a substantial dollar value in their natural state provides an incentive for preservation of marshlands that are in public ownership, but not for preservation of marshlands in private ownership since the owner may receive little or no return, no matter how high is the appraised value to society. The current system, as it now operates, does not work in this case since real estate development of marshlands becomes essentially an irreversible action. It is clear that marshlands must be evaluated as a renewable resource with a value that increases with urban-industrial development.

The time has come to seek means of letting owners of natural resources with high value to society receive a return. The long-term solution is a land-use plan which delimits the amount

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of natural areas that will be necessary to support a future optimum level of urban-industrial development. Then such natural areas can be acquired, or zoned, before the spiral of land speculation raises the market price. The technology of systems ecology is now being developed to the point that an objective compartmentalization between developed and undeveloped environment can be made. Since many coastal wetlands are more productive than adjacent areas, they would generally receive high priority for inclusion into the undeveloped compartment.

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Table 7. Marsh-estuary values as determined by various methods of evaluation.

Basis for evaluation	Annual Return per acre	Income-Capitalization Value per acre (at interest rate 5%)
(1) Commercial & Sports Fisheries	\$ 100	\$2,000
(2) Aquaculture potential (Table 3)		
(a) Moderate oyster culture level	350	7,000
(b) Intensive oyster culture level	900	18,000
(3) Waste treatment (Table 5)		
(a) Secondary	280	5,600
(b) Phosphorous removal ¹	950	19,000
(c) Adjusted tertiary ²	2,500	50,000
(4) Maximum non-competitive summation of values		
(a) 1 + 3c	2,600	52,000
(b) 2b + 3c	3,400	68,000
(5) Total life support value ³	4,150	83,000

¹ Mean of two values shown in Table 5.

² BOD loading (as shown in Table 4) reduced to 3.5 lb/day, a level that reduces O₂ levels about 1 ppm.

³ See text for calculation based on gross primary production.